

ADSORPTION OF CADMIUM (II) IONS BY POLYACRYLONITRILE-BASED ACTIVATED CARBON NANOFIBERS/MAGNESIUM OXIDE AS ITS ADSORBENTS

(Karbon Nano-Gentian Teraktif Berasaskan Poliakrilonitril/Magnesium Oksida Sebagai Penjerap untuk Penjerapan Ion Kadmium (II))

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Received: 21 October 2015; Accepted: 14 June 2016

Abstract

In this work, activated carbon nanofibers (ACNFs) from precursor polyacrylonitrile (PAN) and magnesium oxide (MgO) were prepared via electrospinning process. The morphological properties of the PAN/MgO-based ACNFs were characterized by using Scanning Electron Microscopy (SEM) and the specific surface area (SSA) were investigated using nitrogen adsorption, Brunauer-Emmett-Teller (BET) methods. Moreover, the functional groups were analyzed by using Fourier Transform Infrared (FTIR). Besides that, the sorption study has been conducted in order to determine the adsorption capacity between electrospun ACNFs/MgO, pristine ACNFs and granular activated carbon (GAC) towards cadmium (II) ions. The results showed that the SSA of modified ACNFs ($198.80 \text{ m}^2/\text{g}$) is significantly higher compared to the precursor ACNFs ($15.43 \text{ m}^2/\text{g}$), however the SSA obtained is quite lower compared to the average theoretical value. SEM micrograph of pristine ACNFs depicted more compact nanofibers compared to aligned nanofibers with average diameter of 200-700 nm. Under batch adsorption study, it was found out that Cd(II) removal of both ACNFs and ACNFs/MgO is higher compared to the commercial GAC. It has been demonstrated that the adsorption capacity of both electrospun ACNFs (ACNFs/MgO and pristine ACNFs) is higher when compared to the adsorption capacity of commercial GAC towards Cd (II) ions.

Keywords: polyacrylonitrile, activated carbon nanofiber, magnesium oxide, cadmium (II) ion adsorption

Abstrak

Kajian ini membincangkan tentang proses penyediaan karbon nano-gentian teraktif (ACNFs) berasaskan poliakrilonitril (PAN) dan magnesium oksida (MgO) melalui teknik putaran elektro. Ciri-ciri morfologi bagi ACNFs telah dicirikan menggunakan Mikroskopi Elektron Pengimbasan (SEM) dan luas permukaan kawasan tertentu (SSA) telah disiasat menggunakan kaedah penjerapan nitrogen, Brunauer-Emmett-Teller (BET). Selain itu, kumpulan berfungsi dianalisis dengan menggunakan Inframerah Transformasi Fourier (FTIR). Melalui kajian ini juga, kapasiti penjerapan diantara ACNFs, ACNFs/MgO dan karbon berbutir teraktif (GAC) terhadap ion (II) kadmium telah dijalankan. Hasil kajian menunjukkan bahawa SSA daripada ACNFs yang telah diubah suai ($198.80 \text{ m}^2/\text{g}$) adalah lebih tinggi berbanding dengan pelopor ACNFs ($15.43 \text{ m}^2/\text{g}$), walau bagaimanapun SSA yang diperolehi agak rendah berbanding dengan nilai purata teori. SEM mikrogram bagi pelopor ACNFs menunjukkan nano-gentian yang lebih padat berbanding nano-gentian berstruktur sejajar dengan diameter purata 200-700 nm. Di bawah kajian penjerapan berkumpulan, pembuangan ion (II) kadmium bagi kedua-dua ACNFs dan ACNFs/MgO adalah lebih tinggi berbanding dengan GAC komersial. Hal ini telah membuktikan bahawa kapasiti penjerapan kedua-dua ACNFs yang diperbuat daripada kaedah

putaran elektro (ACNFs/MgO dan ACNFs) terhadap ion (II) kadmium adalah lebih tinggi jika dibandingkan dengan kapasitas penjerapan GAC.

Kata kunci: poliakrilonitril, karbon nano-gentian teraktif, magnesium oksida, penjerapan ion kadmium (II)

Introduction

Nowadays, water contamination by heavy metals remains a major environmental concern as heavy metals persistently present in water subjected by their non-degradable properties [1]. Besides, their tendency to being over-accumulated in biosystem will negatively affect public health and environment despite they play roles in biochemical processes at trace level [2,3,4]. Among the common heavy metals that present in water, cadmium (Cd) has attracted great attention for the studies of its removal as this metal are proved to pose adverse effects on human health even at minimum level [5]. Moreover, the emission of Cd may occur naturally or anthropogenically.

Various methods have been implied upon removal of cadmium laden in water streams such as ion exchange, precipitation, coagulation-flocculation and flotation where each of this method possesses their own limitations more likely to be incomplete removal, generation of secondary waste and inefficient removal efficacies [6,7]. Adsorption is another technique that is frequently applied for removal of heavy metals laden in water. There is wide range of adsorbents available; one good example is activated carbon nanofibers (ACNFs) as it has small inter-fibrous pore size with high porosity and also high specific surface area per unit mass that will lead to high adsorption capacity [8]. One of the challenges to activate the nanofibers is to increase the SSA and capacitance without damaging the graphitic fiber structure, which controls their mechanical properties [9].

Many polymers can be used for preparation of carbon nanofibers such as polyacrylonitrile (PAN), pitch and cellulose. Out of the aforementioned polymers, PAN is an excellent polymer selection because of its outstanding properties which contains a proper functional group(s), adsorbent for metal ion removal, inexpensive, common commercial product and can easily prepared into nanofibers by electrospinning. Preparation of ACNFs with incorporation of nano-scaled metal oxides such as magnesium oxide (MgO) has received great attention as it can solve the major concern of large pores of ACNFs upon removal of heavy metals. Moreover, many previous studies have been conducted on sole MgO due to its large SSA and were proven to be good adsorbent [10]. Thus, in this study the adsorption capacity of prepared ACNFs/MgO were determined and compared with neat PAN-based ACNFs and commercial granular activated carbon (GAC).

Materials and Methods

Materials

Polyacrylonitrile (PAN) with MW of 150, 000 was purchased from Sigma-Aldrich. Dimethylformamide (DMF; ~99.98% purity), magnesium oxides (MgO, ≥99% trace metal basis) and activated carbon (GAC, untreated, granular, 8-20 mesh) were also purchased from Sigma-Aldrich (USA). Stock solution of 1000 mg/L cadmium ions was prepared by using analytical grade cadmium (II) chloride, which was further diluted to the required concentrations before used.

Preparation of PAN nanofibers with dimethylformamide and metal oxides

The dope solution was prepared by adding 1 wt% of MgO into 89 wt% of DMF. The mixture was stirred for at least 5 hours to ensure the MgO dispersed well in the solution before adding up the PAN 10 wt%. The dope was left to stir for 24 hours. The electrospinning process was carried out at fixed electric field of 12kV, the infusion rate is 1 ml/hour and the distance of the tip of needle to the collector is 20 cm, where aluminium foils were used as the collectors.

Activation of PAN-based carbon nanofibers

In stabilization process, the nanofibers were heated up until 275 °C in the furnace (Carbolite) from room temperature with pure air gas flow rate of 2 L/min at heating rate of 2 °C/min with resting time of 30 minutes at 275 °C. The treated nanofibers were further heated up until 600 °C with nitrogen gas flow rate of 0.2 L/min at heating rate of 5 °C /min with 30 minutes of dwelling time and were left for cooling down until room temperature. For

activation process, the carbonized nanofibers were burnt in the tube furnace until 800 °C at rate of 5 °C/min with carbon dioxide gas flow rate of 0.2 L/min and 30 minutes of dwelling time.

Characterization

The surface properties, such as morphology and composition of various adsorbents were determined using scanning electron microscopy (SEM) and the specific surface area were investigated using nitrogen adsorption Brunauer-Emmett-Teller (BET). Fourier Transform Infrared (FTIR) analysis was used to analyze the functional groups appeared in the ACNFs.

Batch adsorption study

Adsorption study of Cd ions was analyzed by investigating the effect of initial concentration upon adsorption rate of Cd. 0.05 g of the prepared samples were placed into conical flask containing 50 mL of Cd solution with initial concentration of 3 ppm respectively. The samples were agitated on rotary shaker with the speed of 180 rpm at 30 °C for 48 hours to reach equilibrium [11,12]. The sample solutions were later filtered and examined by using Atomic Absorption Spectrometer (AAS).

The adsorption capacity of the GAC, ACNFs and ACNFs/MgO and the percentage of removal of metallic ions by adsorbent were calculated by following Equation (1) [13].

$$q_t = \frac{(C_0 - C_t)V}{W} \quad (1)$$

and the removal efficiency was calculated using Equation (2).

$$R(\%) = \frac{(C_0 - C_t)}{C_0} \times 100 \quad (2)$$

where C_0 and C_t are the initial and final concentration of the metallic solution (mg/L), V is the volume of the solution (L), and M is the weight of the adsorbent used (g) [11].

For batch adsorption study, the ACNFs were rinsed with distilled water in order to eliminate any residual solutions. Then, the ACNFs were further dried in oven at room temperature. Desorption of ACNFs have been used to identify the recovery rate and reusability PAN-based ACNFs/MgO. Desorption of metal ions were carried out by immersing the ACNFs in 1M HCl and agitated at 30 °C with speed of 100 rpm for 5 h. The concentration of Cd ions in the solution was analyzed using atomic absorption spectrometer (AAS) and the desorption ratio (D) was calculated as Equation 3 follows:

$$D(\%) = \frac{\text{mg of metal desorbed}}{\text{mg of metal ion adsorbed into ACNFs}} \times 100 \quad (3)$$

Results and Discussion

SEM for morphology characterization

Figure 1 (a and b) showed that both neat PAN-based ACNFs and modified ACNFs (ACNFs/MgO) possessed straight line fibers, bent fibers, rougher and wrinkled fibers due to the shrinkage effect during pyrolysis process. Moreover, in Figure 1(b), the presence of MgO in the ACNFs can be observed by the presence of white spots while the appearance of few beads was believed due to capillary instability [14] during electrospinning.

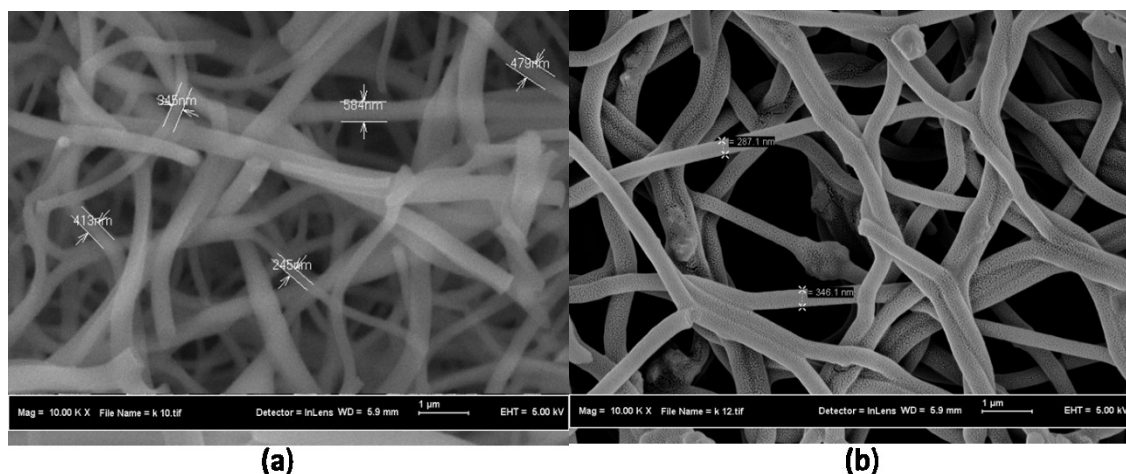


Figure 1. SEM images of (a) ACNFs; and (b) ACNFs/MgO with 10000x magnification

Meanwhile, in comparison of physical observation of SEM micrographs between neat ACNFs and ACNFs/MgO, it can be found that the average diameter of ACNFs without metal oxide is 413.2 ± 128.8 nm (Figure 1a) while for ACNF with MgO is 316.6 ± 41.72 nm as shown in Figure 1(b). As expected, the diameter of ACNFs/MgO is smaller compared to the neat ACNF, as smaller diameter will lead to larger SSA [15,16]. However, even the diameter of the ACNFs obtained not in nano-meter range (below 100 nm), still this study was a success as this result is comparable to other studies on ACNFs.

BET analysis for specific surface area

In this study, the single point BET method was employed to analyze nitrogen adsorption isotherms in order to determine the specific surface area (SSA) of ACNFs and ACNFs/MgO compared to commercial GAC.

Table 1. BET specific surface area of GAC, ACNF, ACNF/MgO

Samples	BET surface area (m ² /g)
GAC	729.82
ACNF	15.43
ACNF/MgO	198.80

From Table 1, the pristine ACNFs show the lowest SSA of 15.43 m²/g while GAC has the highest SSA of 752.60 m²/g. Theoretically, the SSA of both ACNFs and ACNFs/MgO were expected to be much higher compared to the results obtained as tabulated in Table 1. For instance, in a study conducted by Dadvar et al. [17], they have obtained SSA of ACNFs/MgO of 600 m²/g and it was expected in future study, the results for ACNFs/MgO at least comparable or exceeded this value. The low SSA obtained in this study might be due to the incomplete pyrolysis process where the temperature used not high enough to create new micropores instead creating macro- and mesopores that led to low SSA. In our future study, the higher activation temperature with slower heating rate will be used in order to make sure the nanofibers undergo complete pyrolysis stage [18,19].

FTIR for functional group studies

Typically, PAN-based nanofibers will show the existence of CH₂, C≡N, C=O, C–O and C–H bonds but after activation all of these transition compounds were expected to evolve as volatilities and only carbon and hydrogen

atoms remain [20]. As shown in Figure 2, there are six major peaks observed located at 2330.98, 2115.36, 1997.37, 1791.34, 1095.37 and 476.12 cm^{-1} . The peak 2330.98 cm^{-1} is assigned as O-H stretch from strongly hydrogen-bonded- COOH. Peaks 2115.36 and 1997.37 cm^{-1} basically attributed to the strong stretching of C=C groups while peaks 1791.34 cm^{-1} and 1095.37 cm^{-1} indicated as stretching vibrations of C-O groups and C-C stretching, respectively. Last of all, peak 476.12 cm^{-1} were the proofs of the existence of MgO in the ACNFs as the MgO bonds are expected to be presence in the range of 600-450 cm^{-1} [21]. These detected functional groups may play vital role for adsorption of Cd(II).

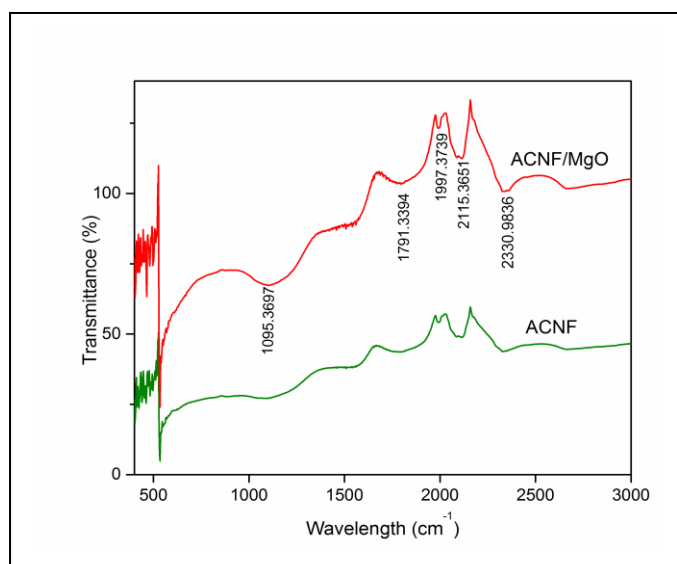


Figure 2. FTIR spectrum of both ACNF and ACNF/MgO

Batch adsorption studies

In this research, the adsorption of heavy metals onto activated carbon nanofibers has been studied using batch-adsorption technique. This study was carried out to examine the adsorption performance of GAC, ACNFs and ACNFs/MgO for the removal of Cd from aqueous solution.

Table 2 shows the percentage removal of cadmium for GAC, ACNFs and ACNFs/MgO it show promising results of 97%, 100% and 100%, respectively. The obtained results revealed that both ACNFs and ACNFs/MgO has good adsorption capacity and effective for the removal of heavy metals from wastewater. Moreover, previous studies have shown that MgO was one of the good adsorbent with high SSA and the incorporation between ACNFs and MgO has proven to be good adsorbent [22] for Cd removal. Based on the results obtained, it can be said that both ACNFs and ACNFs/MgO can be good adsorbent for Cd removal.

Table 2. Removal of cd on GAC, ACNF AND ACNF/MgO

Samples	Removal of Cd (%)
GAC	97
ACNF	100
ACNF/MgO	100

Conclusion

This study investigated the adsorption capabilities of different types of activated carbon on cadmium as a function of adsorbent. Although GAC showed highest SSA compared to both ACNFs, however it was found that pristine and modified ACNFs provided better performance compared to commercial GAC in removal of cadmium from aqueous solution.

Acknowledgement

The authors would like to acknowledge the financial support from the Malaysian Ministry of Education and Universiti Teknologi Malaysia under GUP grants (Q.J130000.2542.05H46 and Q.J130000.2546.12H54) and HICOE grant (R.J090301.7846.4J179). The author would also like to acknowledge the technical and management support from Research Management Centre (RMC), Universiti Teknologi Malaysia.

References

1. Li Y., Wang Y.B., Gou X., Su Y.B. and Wang G. (2006). Risk assessment of heavy metals in soils and vegetables around non-ferrous metals mining and smelting sites, Baiyin, China. *Journal of Environmental Sciences*, 18(6): 1124 – 1134.
2. Kumar R., Isloor A. M., Ismail A. F. and Matsuura T. (2013). Synthesis and characterization of novel water soluble derivative of chitosan as an additive for polysulfone ultrafiltration membrane. *Journal of Membrane Science*, 40: 140 – 147.
3. Meybeck M., Kuusisto A., Makela A. and Malkki R. (1996). Water quality monitoring: a practical guide to the design and implications of freshwater quality studies and monitoring programmes, Water quality. Eds. London, UK: Chapman and Hall.
4. Saeed K., Park S.Y. and Oh T. J. (2012). Preparation of hydrazine-modified polyacrylonitrile nanofibers for the extraction of metal ions from aqueous media. *Journal of Applied Polymer Science*, 121(2): 869 – 873.
5. Vila G., Luger A. and Clodi M. (2012). B-type natriuretic peptide modulates ghrelin, hunger, and satiety in healthy men. *Diabetes*, 61: 2592 – 2596.
6. Barakat M. A. (2011). New trends in removing heavy metal from industrial wastewater. *Arabic Journal of Chemical*, 4: 38 – 41.
7. Fu F. and Wang Q. (2011). Removal of heavy metal ions from wastewater: A review. *Journal of Environment Management*, 92: 20 – 25.
8. Liang Y., Wu D. and Fu R. (2013). Carbon microfibers with hierarchical porous structure from electrospun fiber-like natural biopolymer. *Scientific Reports*, 3: 1119 – 1124.
9. Mao X., Hatton T. and Rutledge G. (2013). A review of electrospun carbon fibers as electrode materials for energy storage. *Current Organic Chemistry*, 17(13): 1390 – 1401.
10. Zahida R., Rabia N., Durr-e-Shahwar, Muhammad Raza S. and Shujat A. (2014). Utilization of magnesium and zinc oxide nano-adsorbents as potential materials for treatment of copper electroplating industry wastewater. *Journal of Environmental Chemical Engineering*, 2(1): 642 – 651.
11. Desta M. B. (2013). Batch sorption experiments: Langmuir and Freundlich isotherm studies for the adsorption of textile metal ions onto teff straw agricultural waste. *Journal of Thermodynamics*, 2013: 1 – 6.
12. Vijayakumar G., Tamilarasan R. and Dharmendirakumar M. (2012). Adsorption, kinetic, equilibrium and thermodynamic studies on the removal of basic dye rhodamine-B from aqueous solution by the use of natural adsorbent perlite. *Journal of Materials and Environmental Science*, 3(1): 157 – 170.
13. Kampalanonwat P. and Supaphol P. (2014). The study of competitive adsorption of heavy metal ions from aqueous solution by aminated polyacrylonitrile nanofibers mat. *Energy Procedia*, 56: 142 – 151.
14. Zussman E., Chen X., Ding W., Calabri L., Dikin D. A., Quintana J. P. and Ruoff R. S. (2005). Mechanical and structural characterization of electrospun PAN-derived carbon nanofibers. *Carbon*, 43: 2175 – 2185.
15. Huang Z. M., Zhang Y. Z., Kotaki M. and Ramakrishna S. (2003). A review on polymer nanofibers by electrospinning and their applications in nanocomposites. *Composites Science and Technology*, 63(15): 2223 – 2253.
16. Drew C., Liu X., Ziegler D., Wang X.Y., Bruno F. F., Whitten J., Samuelson L. A. and Kumar J. (2003). Metal oxide-coated polymer nanofibers. *American Chemical Society*, 3(2): 143 – 147.
17. Dadvar S., Tavanai H. and Morshed M. (2012). Effect of embedding MgO and Al₂O₃ nanoparticles in the precursor on the pore characteristics of PAN based activated carbon nanofibers. *Journal of Analytical and Applied Pyrolysis*, 98: 98 – 105.

18. Abdullah A. H., Kassim A., Zainal Z., Hussein M. Z., Kuang D., Ahmad F. and Wooi O.S. (2001). Preparation and characterization of activated carbon from gelam wood bark. *Malaysian Journal of Analytical Sciences*, 7(1): 65 – 68.
19. Bonso J. S., Kalaw G. D. and Ferraris J. P. (2014). High surface area carbon nanofibers derived from electrospun PIM-1 for energy storage applications. *Journal of Materials Chemistry A*, 2: 418 – 424.
20. Arshad S. N., Naraghi M. and Chasiotis I. (2011). Strong carbon nanofibers from electrospun polyacrylonitrile, *Carbon*, 49: 1710 – 1719.
21. Hofmeister M., Keppel E. and Speck K. (2003). Adsorption and reflection infrared spectra of MgO and other diatomic compounds. *Monthly Notices of the Royal Astronomical Society*, 345(1): 16 – 38.
22. Rajagopalan S., Koper O., Decker S. and Klabunde K. J. (2002). Nanocrystalline metal oxides as destructive adsorbents for organophosphorus compounds at ambient temperatures. *Chemistry-A European Journal*, 8(11): 2602 – 2607.